

LEACHATE GENERATION, COLLECTION, AND TREATMENT

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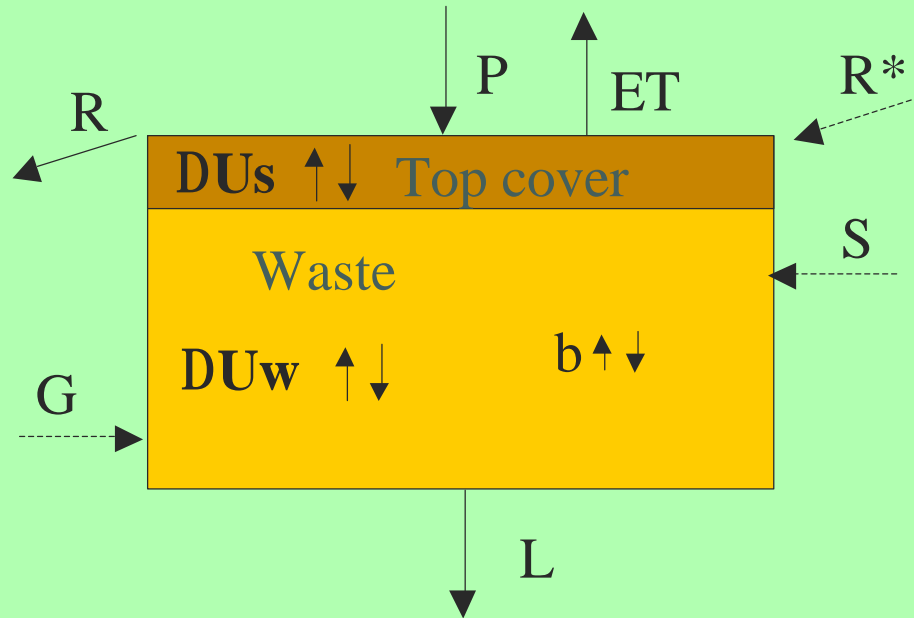


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I M A G E

LEACHATE PRODUCTION





$$L = P + S + G + R^* - R + \Delta U_s + \Delta U_w - ET + b$$

L = Leachate generated

P = Precipitation (actually plus recirculated leachate and surface input)

S, G = infiltration from surface water or groundwater

ΔU_s = Change in moisture storage in top cover

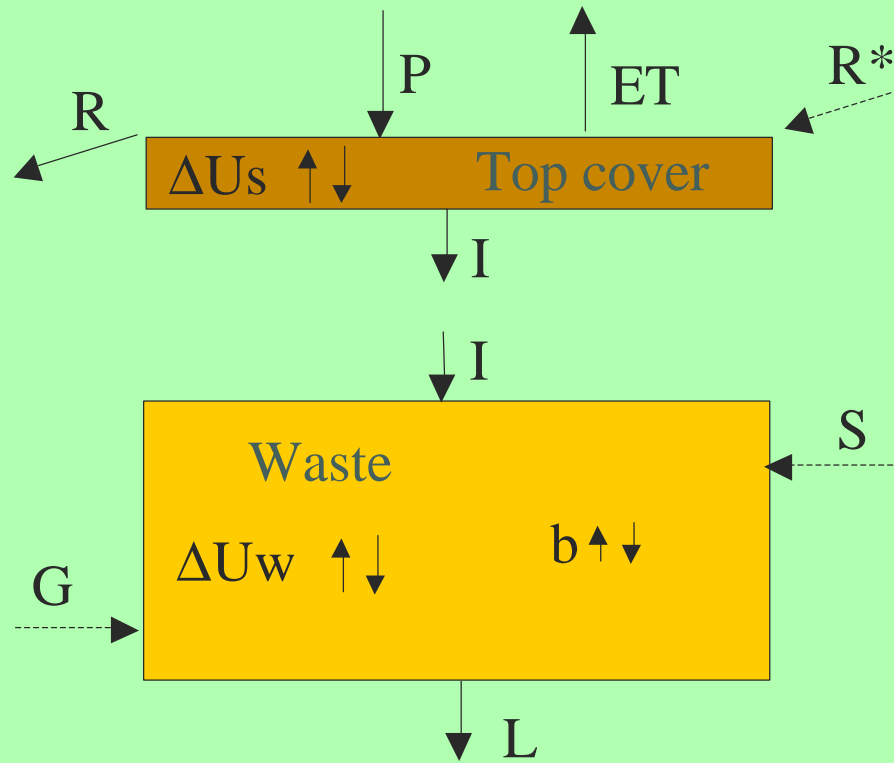
ET = Actual evapotranspiration

R, R* = Surface runoff

ΔU_w = Change in moisture content of refuse

b = biochemical water production or consumption





$$I = P + R^* - R + \Delta U_s - ET$$



SURFACE RUNOFF

$$R = C \cdot P$$

R = surface runoff (mm/d)

C = runoff coefficient

P = rainfall (mm/d)

$$C = a \cdot b_i$$

a – depends on the presence of the final cover, on the kind of materials used and on the slope.

b – depends on soil moisture content in the different months



RUNOFF COEFFICIENT $C = a \cdot b_i$

Empirical values for “a”

| Landfill | type of soil | slope | | |
|--------------|--------------|-----------|-----------|-----------|
| | | <5% | 5-10% | >10% |
| closed | sandy | 0,05-0,10 | 0,10-0,15 | 0,15-0,20 |
| | clayey | 0,13-0,17 | 0,18-0,22 | 0,25-0,35 |
| in operation | sandy | 0,08-0,13 | 0,13-0,18 | 0,18-0,25 |
| | clayey | 0,16-0,20 | 0,21-0,25 | 0,27-0,38 |

Empirical values of “b” for Italy

| Month | (b _i) | Month | (b _i) |
|----------|-------------------|-----------|-------------------|
| January | 1,60 | July | 0,29 |
| February | 1,80 | August | 0,29 |
| March | 1,43 | September | 0,46 |
| April | 0,97 | October | 1,20 |
| May | 0,89 | November | 1,40 |
| June | 0,37 | December | 1,60 |



Evapotranspiration (ET)

- Potential ET (ET_p): Maximal ET from surface covered with a homogeneous, green crop with optimal water supply
- Governing factors:
 - Meteorological factors: Wind, Temperature, Relative humidity
 - Soil and plant factors: Type/state of crop, Soil type

Actual evapotranspiration

$$ET = ET_p \cdot U/FC$$

U actual moisture content
FC field capacity

POTENTIAL EVAPOTRANSPIRATION

Thorntwaite Formula:

$$PE_i = 16 \left(\frac{10T_i}{I_T} \right)^a \cdot C_i$$

- PE_i = potential evapotranspiration of the i-month (mm/month)

T_i = monthly average temperature (°C)

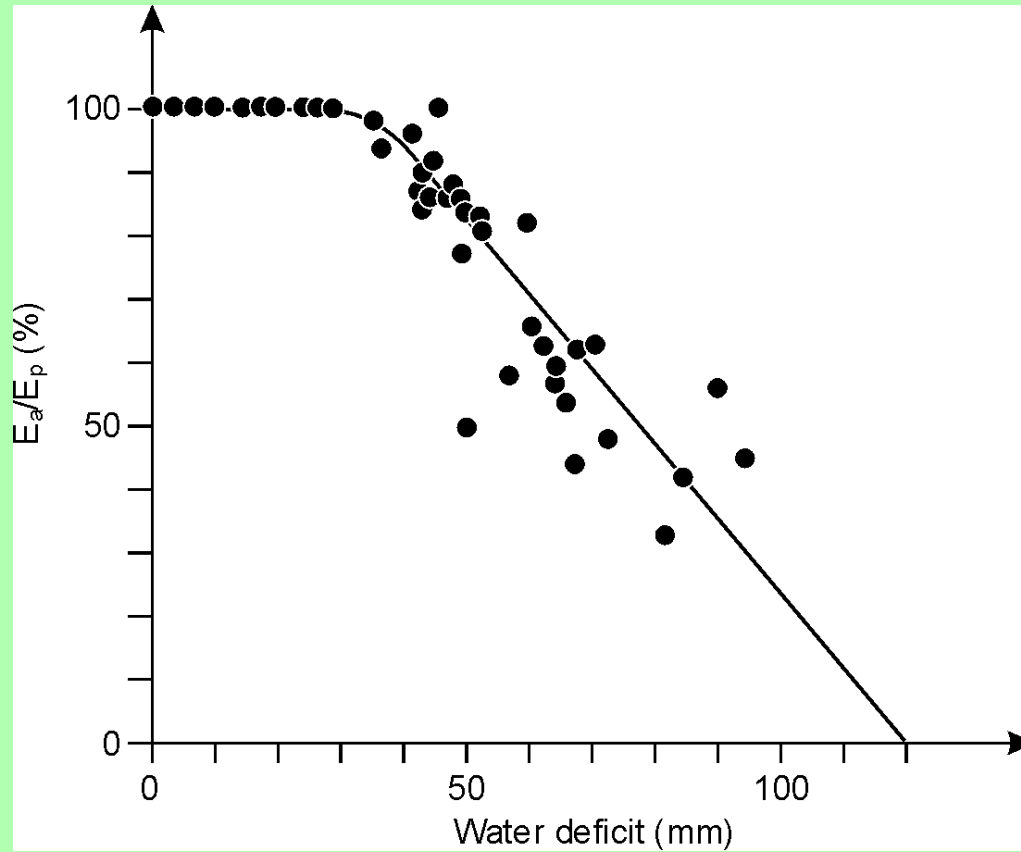
$$I_T = \sum_{i=1}^{12} \left(\frac{T_i}{5} \right)^{1,514} = \text{annual thermal index}$$

$$a = 6,75 \cdot 10^{-7} I_T^3 - 7,71 \cdot 10^{-5} I_T^2 + 1,79 \cdot 10^{-2} I_T + 0,49239$$

C_i = depends on hours of sunlight and on latitude



Actual evapotranspiration



Estimation of Leachate Volume

A rough estimation of leachate production may be given as a percentage of rainfall, as a function of waste density in landfill

low compacted landfill: 25 - 50 % of rainfall

high compacted landfill: 15 - 25 % of rainfall

As a consequence, for an average precipitation of 700 mm/y the leachate production expected is:

- low compacted landfill : $5 - 10 \text{ m}^3 / (\text{ha}\cdot\text{d})$
- high compacted landfill : $4 - 5 \text{ m}^3 / (\text{ha}\cdot\text{d})$



Leachate production and composition

Leachate production depends on:

Climate

Morphology of the underground

Composition and quality of waste

Landfill operation (daily cover, final closure)

Leachate Composition depends on

Composition and quality of waste

Emplacement technology

Emplacement speed

Water balance



Measures to be taken

Minimisation of leachate production (???)

Leachate collection at landfill base

Leachate discharge from landfill body to treatment plant

Minimization of the leachate head in the landfill body

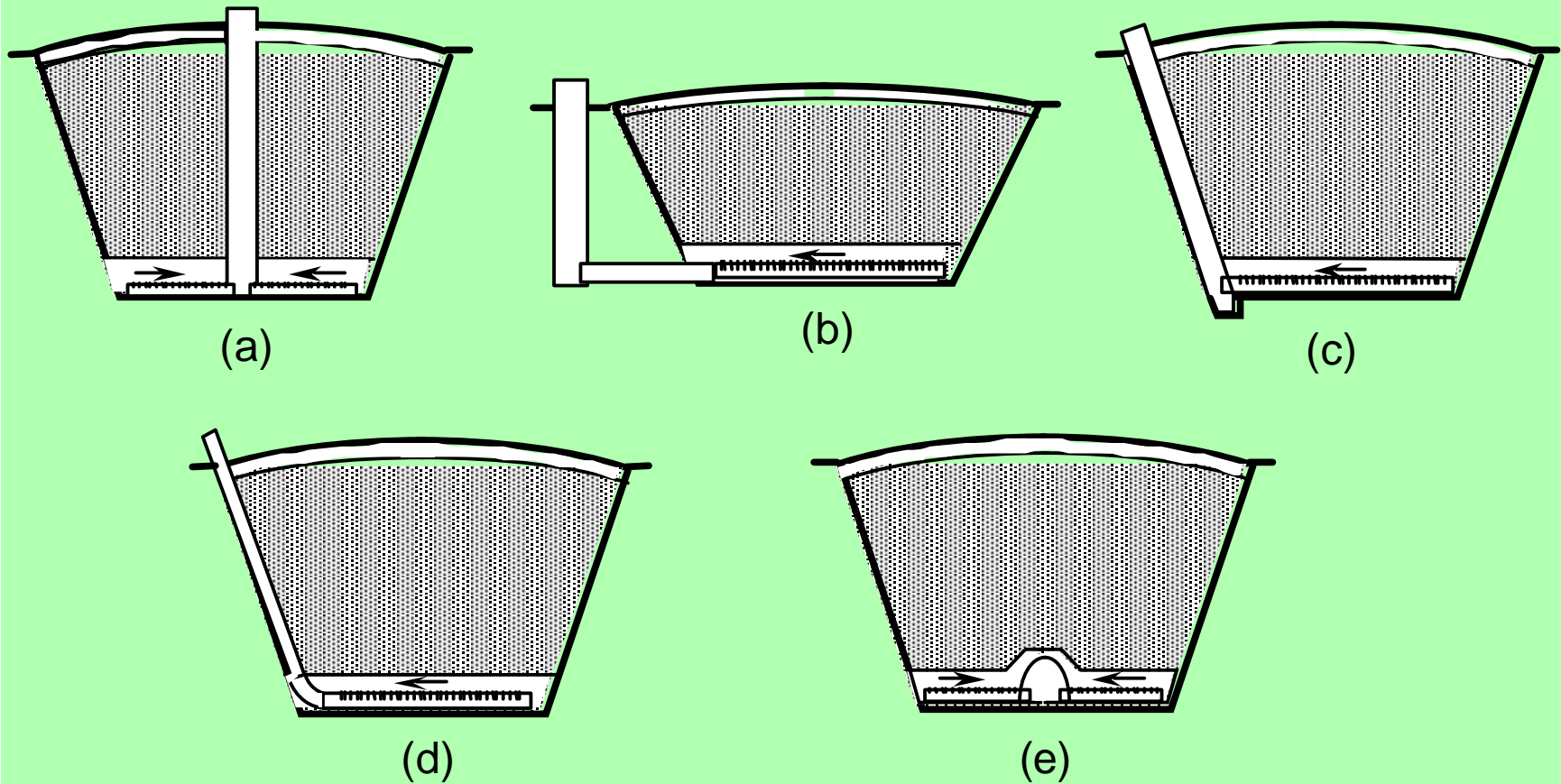
Ranke: „Sickerwassersammlung und -ableitung“ in MHB, KZ 4545, 1998



LEACHATE COLLECTION

- Design strategies are addressed to minimize the transport of contaminants through the barriers to the environment.
- Drainage and collection systems are essential components in a containment landfill and can be considered as a barrier.
- When efficient they prevent leachate buildup and consequently decrease potential leakage to groundwater through the low permeability bottom liner.





Different possible configurations for leachate extraction wells: a = Central internal shaft; b= Lateral out-site well; c = Lateral slope internal shaft; d = Draining pipe - shaft; e = Tunnel

Engineering options to be employed for leachate collection

a) Drainage

- saw tooth configuration for the landfill bottom
- good longitudinal slope ($> 2\%$)
- minimum cross slope of 1% (23% is desirable)
- use, at least for MSW landfills, of only granular material (gravel)
- high porosity of drainage layer
- thickness of the drainage layer > 50 cm
- clean gravel, possibly round shaped
- high grain size which is compatible with filter stability
- aerial distribution of the drainage. The filter should cover the entire area of the landfill bottom and slope
- split gravel trenches suitable for a drainage layer of fine material should be placed at a distance of 15 - 20 m and should have a minimum width of 2m
- avoidance of the use of any filter material (such as geotextiles) to "protect" the drainage from clogging They actually protect the filter too well!

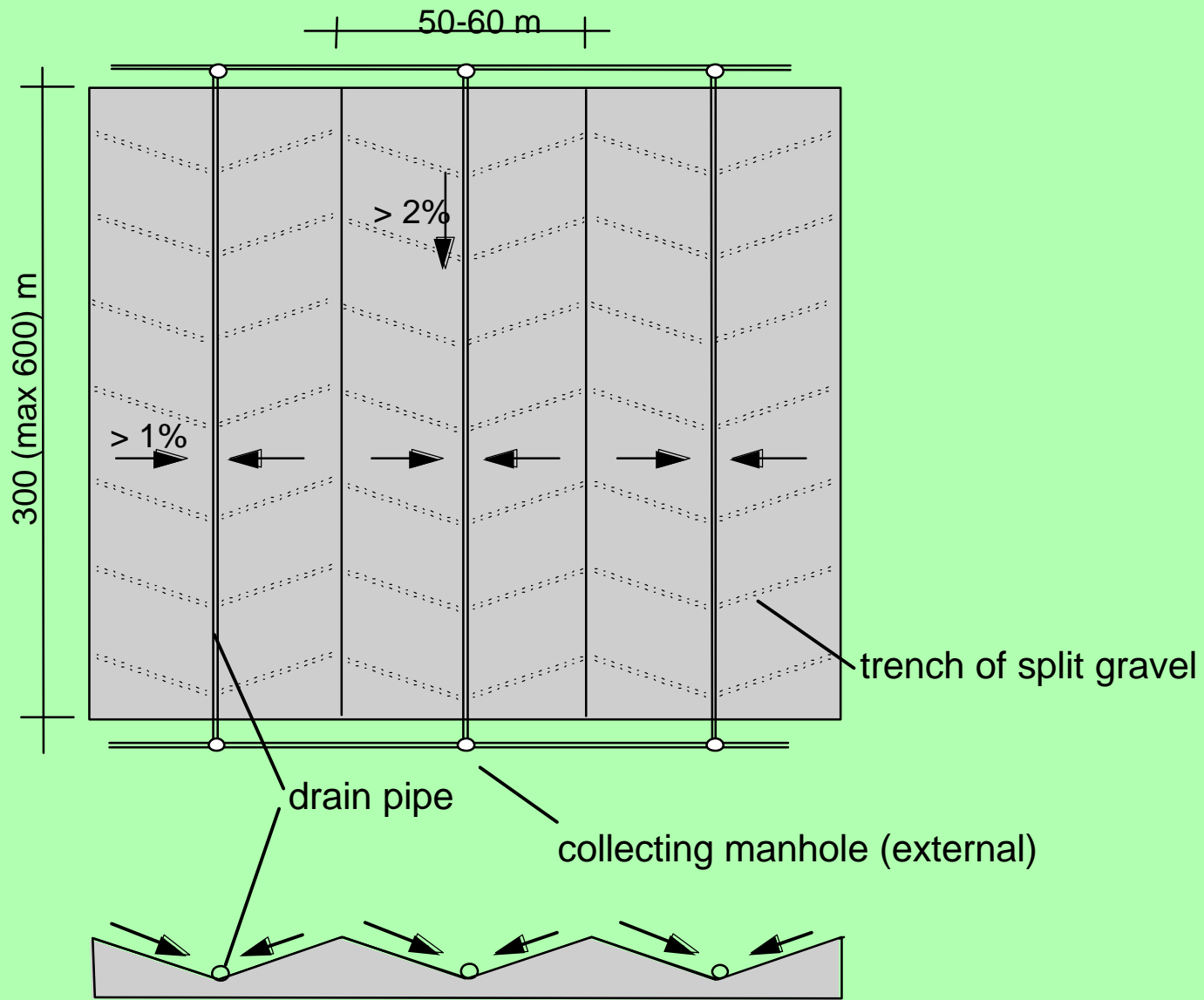


Engineering options to be employed for leachate collection

a) Drain pipes

- short space between drain pipes (50 - 60 m)
- parallel straight line layout of drain pipes
- pipes should be accessible from outside the waste deposit
- pipe diameter should be larger than 200 mm
- pipeline should be designed according to a good pressure distribution under the given conditions
- reinforced drain pipe should be consequently adopted
- drain pipes should be controlled by a videocamera immediately after implacement of the first waste lift. Mechanical failure caused by compaction can be easily repaired at this stage.
- regular flushing of drain pipes should be carried out as soon as possible as clogging deposits can be removed before concreting
- process enhancement, as discussed earlier, has a positive effect on the efficiency of drainage layer as the methanogenic leachate proved to be less clogging (Ramke,1989; McBean *et al.*,1993).





LEACHATE QUALITY



Leachate and gas evolution in a landfill

(Kristensen & Kjeldsen (1989))

Phase I:

Aerobic phase

Phase II:

Acid phase

Phase III:

Intermediate
methanogenic phase

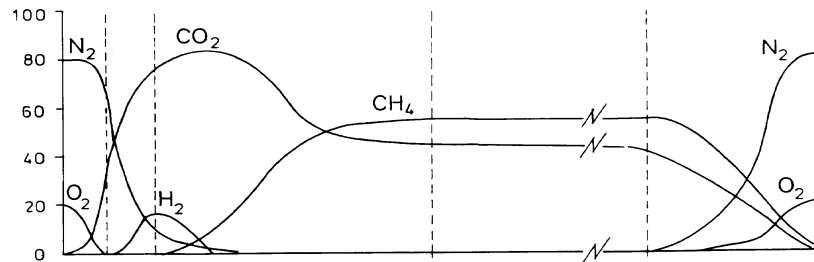
Phase IV:

Stabilized
methanogenic phase

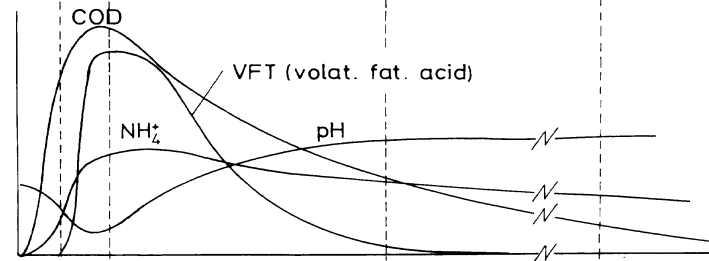
Phase V:

Final aerobic phase

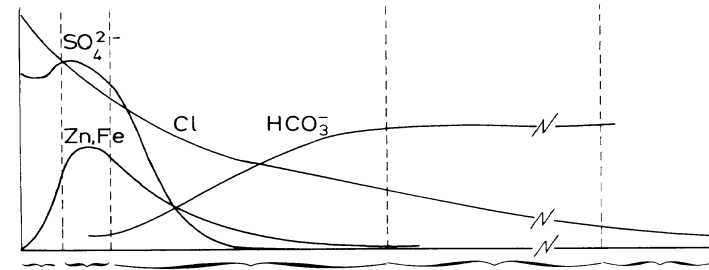
Gas composition, vol %



Leachate



Leachate



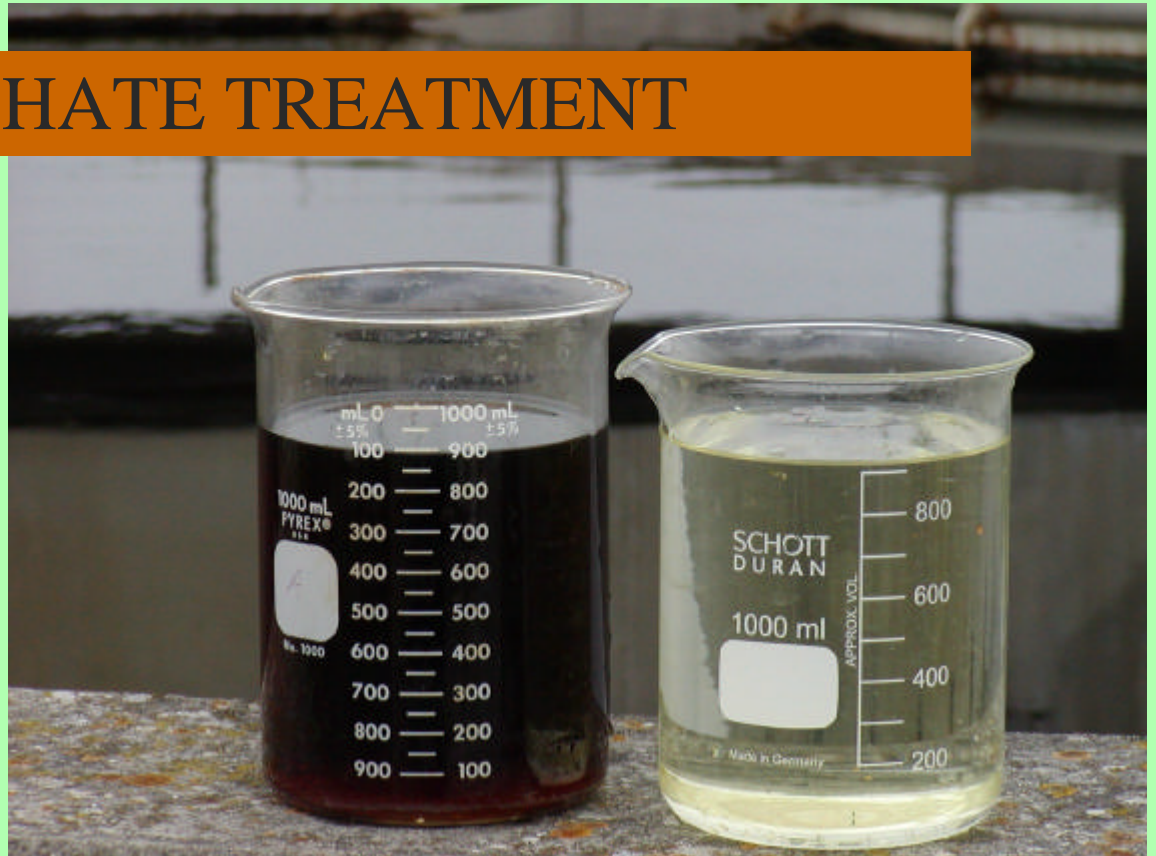
Phase I II III IV V

Constituents in leachate (Ehrig 1990)

| Parameter | Unit | Leachate from MSW landfills | |
|----------------------------------|------|-----------------------------|--------------------|
| | | Acid phase | Methanogenic phase |
| | | Range | Range |
| pH-value | - | 4,5 - 7 | 7,5 - 9 |
| COD | mg/l | 6.000 - 60.000 | 500 - 4.500 |
| BOD ₅ | mg/l | 4.000 - 40.000 | 20 - 550 |
| TOC | mg/l | 1.500 - 25.000 | 200 - 5.000 |
| AOX | µg/l | 540 - 3.450 | 524 - 2.010 |
| org. N ¹⁾ | mg/l | 10 - 4.250 | 10 - 4.250 |
| NH ₄ -N ¹⁾ | mg/l | 30 - 3.000 | 30 - 3.000 |
| TKN ¹⁾ | mg/l | 40 - 3.425 | 40 - 3.425 |
| NO ₂ -N ¹⁾ | mg/l | 0 - 25 | 0 - 25 0 |
| NO ₃ -N ¹⁾ | mg/l | 0,1 - 50 | 0,1 - 50 |
| SO ₄ | mg/l | 70 - 1.750 | 10 - 420 |
| Cl | mg/l | 100 - 5.000 | 100 - 5.000 |



LEACHATE TREATMENT



LEACHATE TREATMENT

- Leachate treatment is a complex task because:
 - every landfill have different characteristics
 - flow rates and pollutant concentrations are variable with time and with the position
 - concentrations of some compounds might be very high
 - the nature of the contaminants may prevent the application of biological processes
- Treatment options:
 - on site treatment plant
 - combined treatment with sewage
 - leachate recirculation



LEACHATE TREATMENT

Treatment Methods

Biological treatments (aerobic, anaerobic)

Adsorption

Chemical oxidation

Membrane technology

Evaporation

Desiccation

Stripping

Flocculation / precipitation

Filtration

Sedimentation



Biological treatment



Biological treatment

- Objectives:
 - reduction of organic load
 - nitrification and denitrification
- Main control parameters: COD, BOD, ammonia nitrogen, heavy metals
- Need of flexible solutions due the variations of leachate quality and quantity.
- Less expensive than chemico-physical treatment processes, however the law limits cannot be reached when used as unique step
- No residues production except the biomass



ANAEROBIC TREATMENT

- Usually applied for young leachate and as pretreatment before aerobic treatment, as less sensitive to flow rate and load variation.
- Advantages:
 - methane production;
 - no need for aeration, limited energy required;
 - low sludge production
- Disadvantages:
 - sensitivity to pH, metals, phenols...
 - BOD, COD, and N-NH₄ residual concentrations still too high
- The efficiency depends on volumetric load and temperature



Some important reactions for four groups of bacteria involved in anaerobic waste degradation

| | |
|------------------------------------|---------------------------------|
| Fermentative processes | |
| $C_6H_{12}O_6 + 2H_2O$ | $2CH_3COOH + H_2 + 2CO_2$ |
| $C_6H_{12}O_6$ | $CH_3C_2H_4COOH + 2H_2 + 2CO_2$ |
| $C_6H_{12}O_6$ | $2CH_3CH_2OH + 2CO_2$ |
| Acetogenic processes | |
| $CH_3CH_2COOH + 2H_2O$ | $CH_3COOH + 3H_2 + CO_2$ |
| $CH_3C_2H_4COOH + 2H_2O$ | $2CH_3COOH + 2H_2$ |
| $CH_3CH_2OH + H_2O$ | $CH_3COOH + 2H_2$ |
| $C_6H_5COOH + 4H_2O$ | $CH_3COOH + H_2$ |
| Methanogenic processes | |
| $4H_2 + CO_2$ | $CH_4 + 2H_2O$ |
| CH_3COOH | $CH_4 + CO_2$ |
| $HCOOH + 3H_2$ | $CH_4 + 2H_2O$ |
| $CH_3OH + H_2$ | $CH_4 + H_2O$ |
| Sulphate reducing processes | |
| $4H_2 + SO_4^{2-} + H^+$ | $HS^- + 4H_2O$ |
| $CH_3COOH + SO_4^{2-}$ | $CO_2 + HS^- + HCO_3^- + H_2O$ |
| $CH_3C_2H_4COOH + SO_4^{2-} + H^+$ | $4CH_3COOH + HS^-$ |

$HCOOH$: formic acid,
 CH_3COOH : acetic acid,
 CH_3CH_2COOH : propionic acid,
 $CH_3C_2H_4COOH$: butyric acid,
 $C_6H_{12}O_6$: glucose,
 CH_3OH : methanol,
 CH_3CH_2OH : ethanol,
 C_6H_5COOH : benzoic acid,
 CH_4 : methane,
 CO_2 : carbon dioxide,
 H_2 : hydrogen,
 SO_4^{2-} : sulphate,
 HS^- : hydrogen sulphide,
 HCO_3^- : hydrogen carbonate,
 H^+ : proton,
 H_2O : water.

(Kristensen & Kjeldsen (1989))



Anaerobic treatment

•DIGESTORS:

Closed reactors continuously mixed by means of injection of the produced biogas; the temperature is kept above 20°C.

Higher efficiency in comparison to lagoons, but higher costs.

The efficiency of BOD and COD removal is about 80-90% in case of volumetric loads of about 1 kg COD/m³d and for T=25-30 °C.

The precipitation of metals as sulphides may reach an efficiency of 90%.

Residence time:

- Digestors: 12-15 d
- Lagoons: 20-40 d



Anaerobic treatment

- UASB Reactors:

Leachate is pumped from the bottom of the reactor, through a filter of porous media with high specific surface, where the biomass is attached.

Advantages: higher biomass concentration and thus treatment of leachate with higher organic volumetric load (up to 4 kgCOD/m³-d), lower retention times, lower volumes and higher biogas production (about 6 liters of biogas per liter of leachate).

Disadvantages: higher costs for plant installation, partly offset by energy recovery from the biogas.



APPLICATION OF REMOVAL KINETICS

COD removal was found to follow a first order rate equation such that removal rate was independent of the initial COD concentration.

$$-d[\text{COD}]/dt = K[\text{COD}]$$

where

[COD] = initial COD concentration (mg/litre)

t = retention time (days)

K = reaction constant (day⁻¹)

Integrating eqn (1) between the initial time (t₀) and time t yields

$$\ln[\text{COD}]_t / [\text{COD}]_0 = -Kt$$

Using the relationships described above, an expression can be derived linking the mean operating temperature and the target percentage COD removal (η COD) to the volume of a suitable anaerobic lagoon

$$\eta\text{COD} = [\text{COD}]_0 - [\text{COD}]_t / [\text{COD}]_0$$

$$\eta\text{COD} = 1 - e^{-Kt}$$

If Q equals the daily leachate flow (m³/day) and V is the volume of the anaerobic lagoon, then rearranging and substituting V/Q for t:

$$V = Q \ln (1 - \eta\text{COD}) / K$$

Experimental values

$$k_{25^\circ} = 0,0317 \text{ d}^{-1}$$

$$k_{10^\circ} = 0,0083 \text{ d}^{-1}$$

$$k_{4^\circ} = 0,0012 \text{ d}^{-1}$$

$$V = Q \ln(1 - \eta_{\text{COD}})/K$$

Mean daily flow from a hypothetical landfill receiving a mean annual rainfall of about 700 mm/year has been estimated to be approximately 3 m³/ha day (Ehrig). Extrapolating this figure to a 10-ha site, daily flow would be around 30 m³. A relationship between leachate temperature and lagoon volume for a range of treatment efficiencies is shown.

Estimates of lagoon volume for a 30 m³/day leachate flow under different temperature conditions for a range of treatment efficiencies

| Target COD removal (%) | Lagoon volume (m ³ x 10 ³) | | |
|------------------------|---|-------|-------|
| | 4 °C | 10 °C | 25 °C |
| 10 | 2.63 | 0.38 | 0.01 |
| 25 | 7.19 | 1.04 | 0.27 |
| 50 | 17.3 | 2.51 | 0.66 |
| 75 | 34.7 | 5.01 | 1.31 |
| 90 | 57.6 | 8.32 | 2.18 |

AEROBIC TREATMENT

- Suitable for the treatment of young leachate.
- The processes are more rapid than anaerobic ones but more expensive, due to the need of providing aeration.
- Less sensitive than anaerobic ones to the presence of inhibitors such as heavy metals, phenols, sulphides, ammonia...
- Much sensitive to load fluctuations.
- Removal efficiency for COD e BOD: 98-99%; ammonia nitrogen: 90% when NH_4 in the leachate is below 100-200 mg/l and 80% when above 500 mg/l.



Aerobic degradation reactions

aerobic degradation



Nitrification



Denitrification



Albers, Ehrig, Mennerich : Sickerwasserreinigung. Müll-Handbuch. 4588. Lfg. 1/91. Erich Schmidt Verlag, Berlin





Aerated Lagoons



AEROBIC TREATMENT

Aerated Lagoons

Depth: 2-6 m

Aerobic conditions obtained by means of turbines.

No recirculation, the volume has to guarantee retention times high enough for biomass growth.

Very high efficiency for $BOD/COD > 0,4$; $T > 20^{\circ}C$; 20-30 day retention time; volumetric load 1-1,5 $kgCOD/m^3d$

The efficiency depends on temperature, on volumetric load and retention time (t).

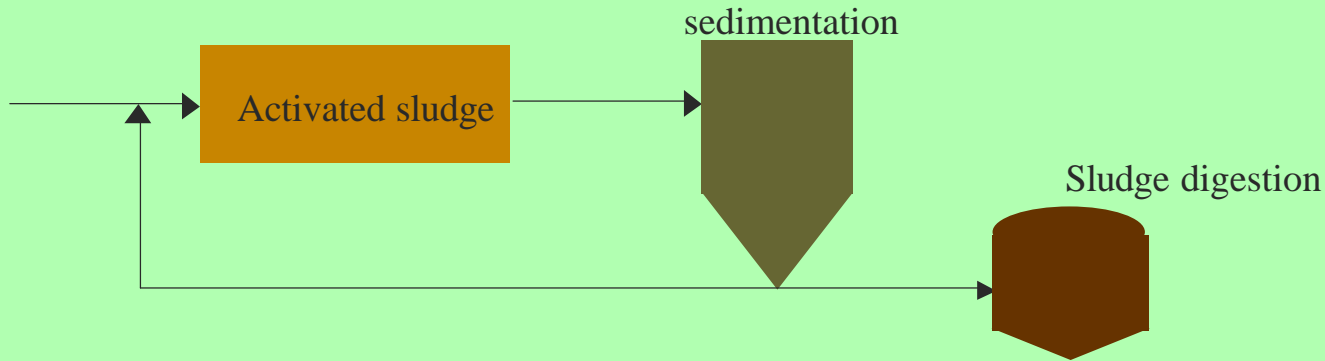
High efficiency for nitrification for retention times > 20 d

Metal removal due to precipitation as hydroxides.



Activated sludge processes

Detention time considerably shorter than in aerated lagoons as the sludge content can be controlled and is 3-5 times higher. This is achieved by installing a settling tank and recirculating the sludge back into the activated sludge tank.



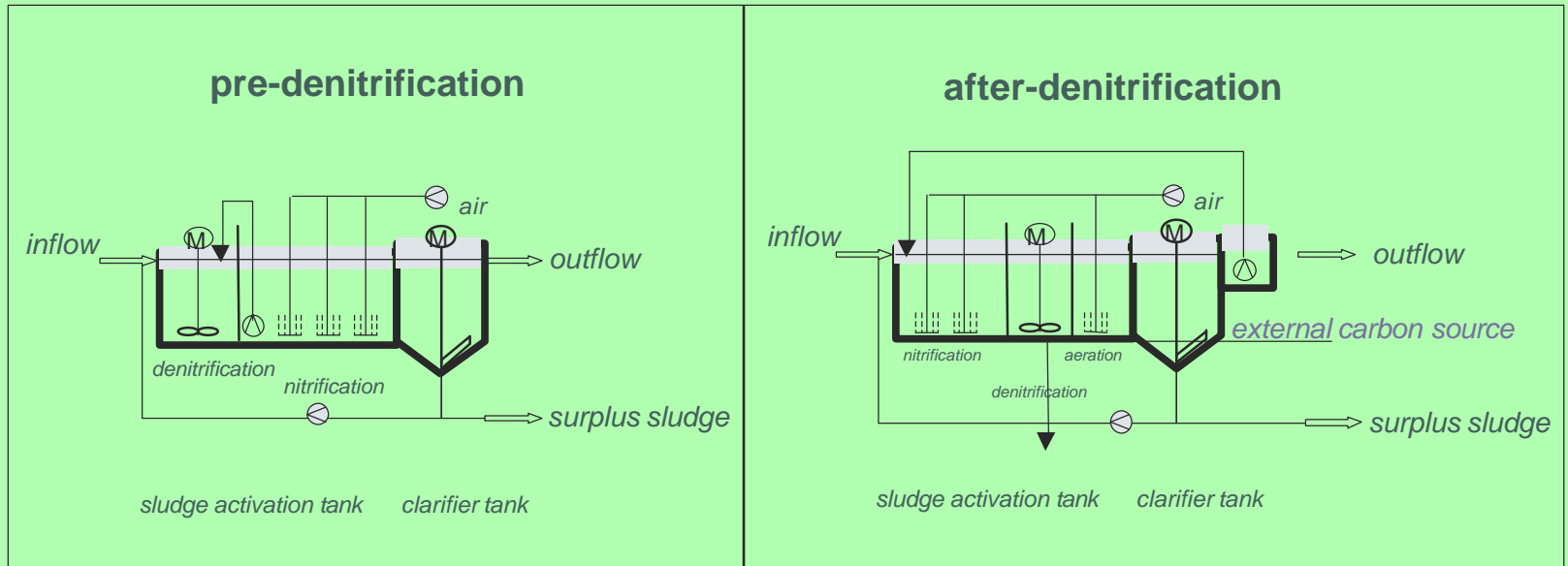
$F/M = 0.02 - 0.05 \text{ kg BOD}_5/\text{kgMLSS}\cdot\text{d}$

Volumetric Load: $0.2 - 0.6 \text{ kg BOD}_5/\text{m}^3\cdot\text{d}$

Effluent: $\text{BOD}_5 < 25 \text{ mg/l}$

Activated sludge processes

NITRIFICATION OF AMMONIA



In general, complete nitrification can be observed at N-loading rates lower than 0,03 kgN/kg MLSS·d. A denitrification step is necessary also.

Aerobic treatment

Rotating biological contactors (RBC):

This process differs from the activated sludge process in so far as the bacteria are attached to the rotating contactors. The air supply takes place naturally, i.e. the rotating contactor is partly in the air and partly in the water.

Low amount of energy required.

Main advantage: high efficiency (up to 95%) for nitrification.

In order to avoid toxic nitrite concentrations, the nitrogen loading should not exceed $2\text{g N/m}^2\cdot\text{d}$

Some authors report that about 95% of ammonium concentrations are oxidized also when high loading rates are present ($> 10\text{g N/m}^2\cdot\text{d}$).



Aerobic treatment

Trickling filters

In trickling filters air vents from the bottom to the top through a filter of porous media with high specific surface, where the biomass is attached. Leachate is sprinkled on the surface and the treatment occurs during the percolation through the filter material. When treating highly organically-polluted leachate, clogging by inorganic precipitates or produced biomass may occur.

Good efficiency of nitrification (up to 95%), proved for leachate with the following characteristics:

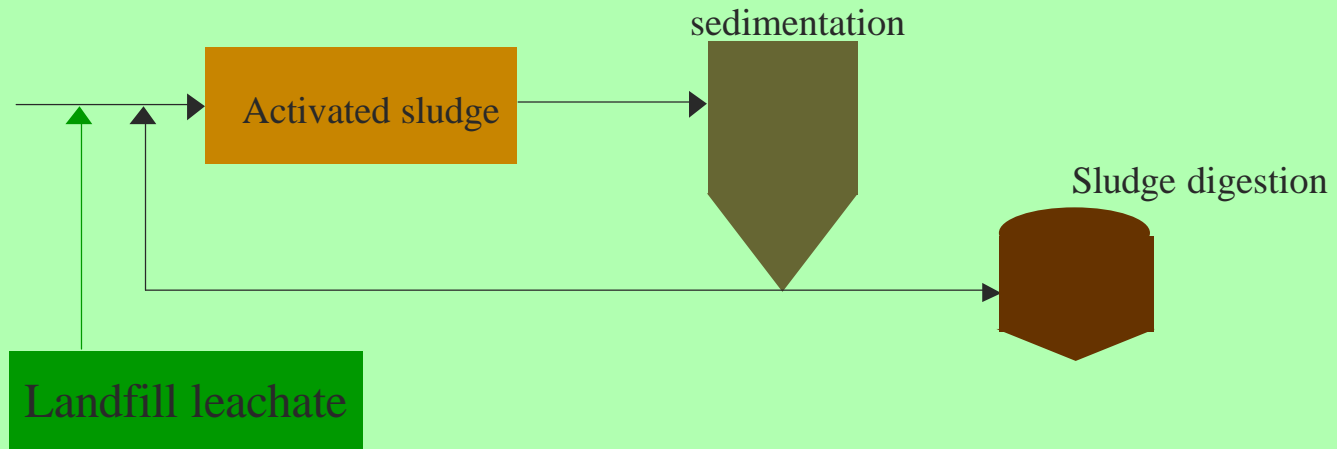
Ammonia nitrogen: 200-600 mg/l;

BOD: 80-250 mg/l;

COD: 850-1350 mg/l;



CO-TREATMENT IN DOMESTIC SEWAGE FACILITY



Problems:

Organic load

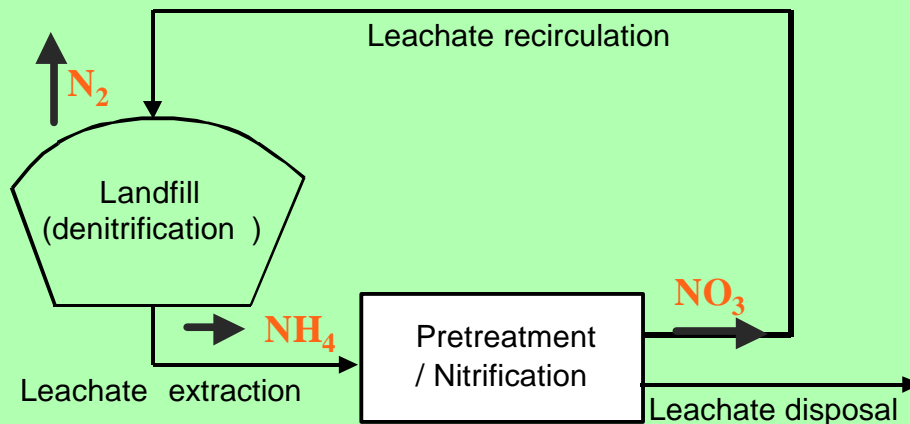
Dilutions of metals and AOX

Nitrogen and Phosphorous



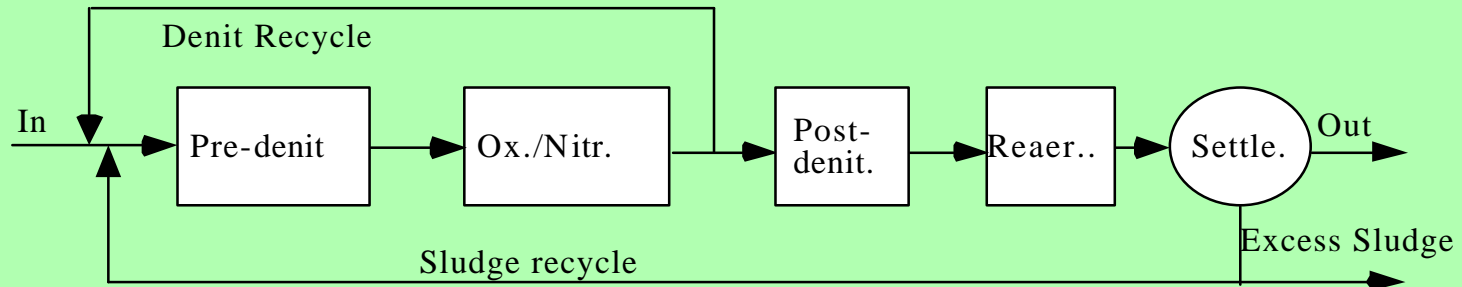
LEACHATE RECIRCULATION

- Reduction of the amount of leachate to be treated by increasing evapotranspiration
- Reduction of the organic content of leachate and of the cost for treatment
- Enhancement of degradation processes by increasing water content and supply and distribution of nutrients and biomass
- Dilution of locally high concentrations of inhibitors
- Possible recirculation after nitrification



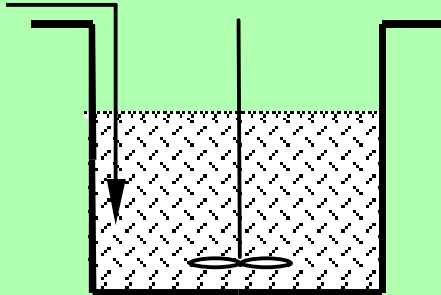
SBR

SBRs perform in a single reactor, during temporised cycles, the same reactions that the continuous flow treatment trains do in different reactors

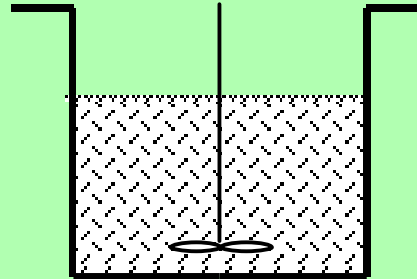


SBR phases

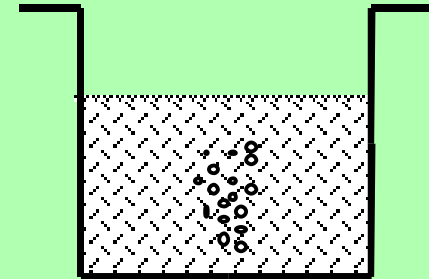
Phase 1: mixed feed



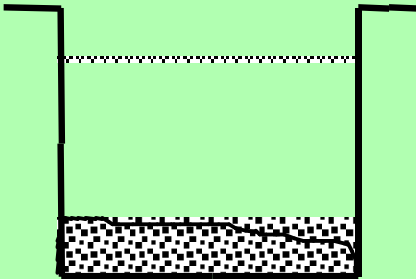
Phase 2a: mixed
(anoxic/anaerobic) reaction



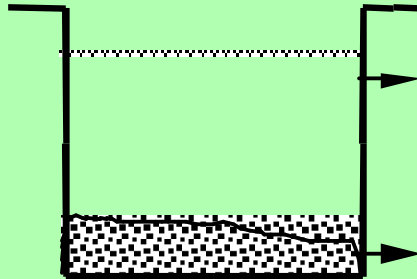
Phase 2b: mixed
aerated reaction



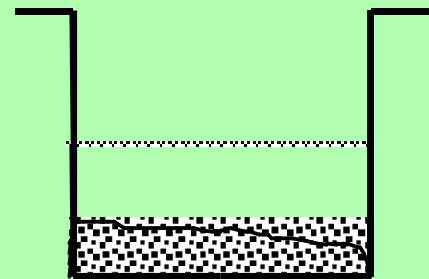
Phase 3: settling phase



Phase 4: sludge/effluent
withdrawal



Phase 5: idle





Plant under construction, showing the twin row leachate balancing tanks, three large SBR tanks, and effluent balance tank and reed beds top right (Robinson, Sardinia 2003)





Effluent balance tank, and recently planted reed beds (Robinson, Proceedings Sardinia 2003)



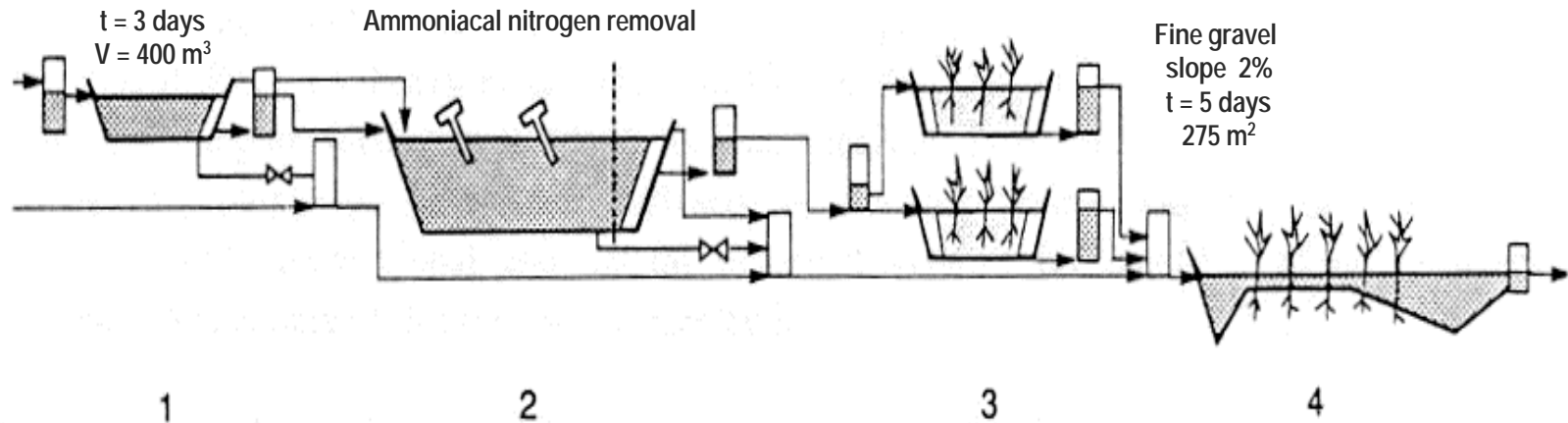
PLANT PERFORMANCE

| Determinand | Influent leachate | Effluent |
|------------------|-------------------|----------|
| COD | 4730 – 5990 | 1010 |
| BOD ₅ | 537 – 688 | <1 |
| ammoniacal-N | 1240 - 1460 | 1.5 |

The effluent from the plant is polished by passage through a series of terraced reed beds, a wholly natural process in which the reed plant rhizomes provide additional treatment to high standards, before final effluent is discharged into the Mersey river at typical rates of 10 – 20 cubic metres per hour (Robinson, Sardinia 2003).



Leachate treatment plant of Esval



1 - anaerobic lagoon

2 - aerobic lagoon

3 - two parallel CWs wit subsurface flux

4 - terminal surface flux CW

Efficiency for BOD_5 and NH_3 is 60-90% depending on season (Maelun - 1995)



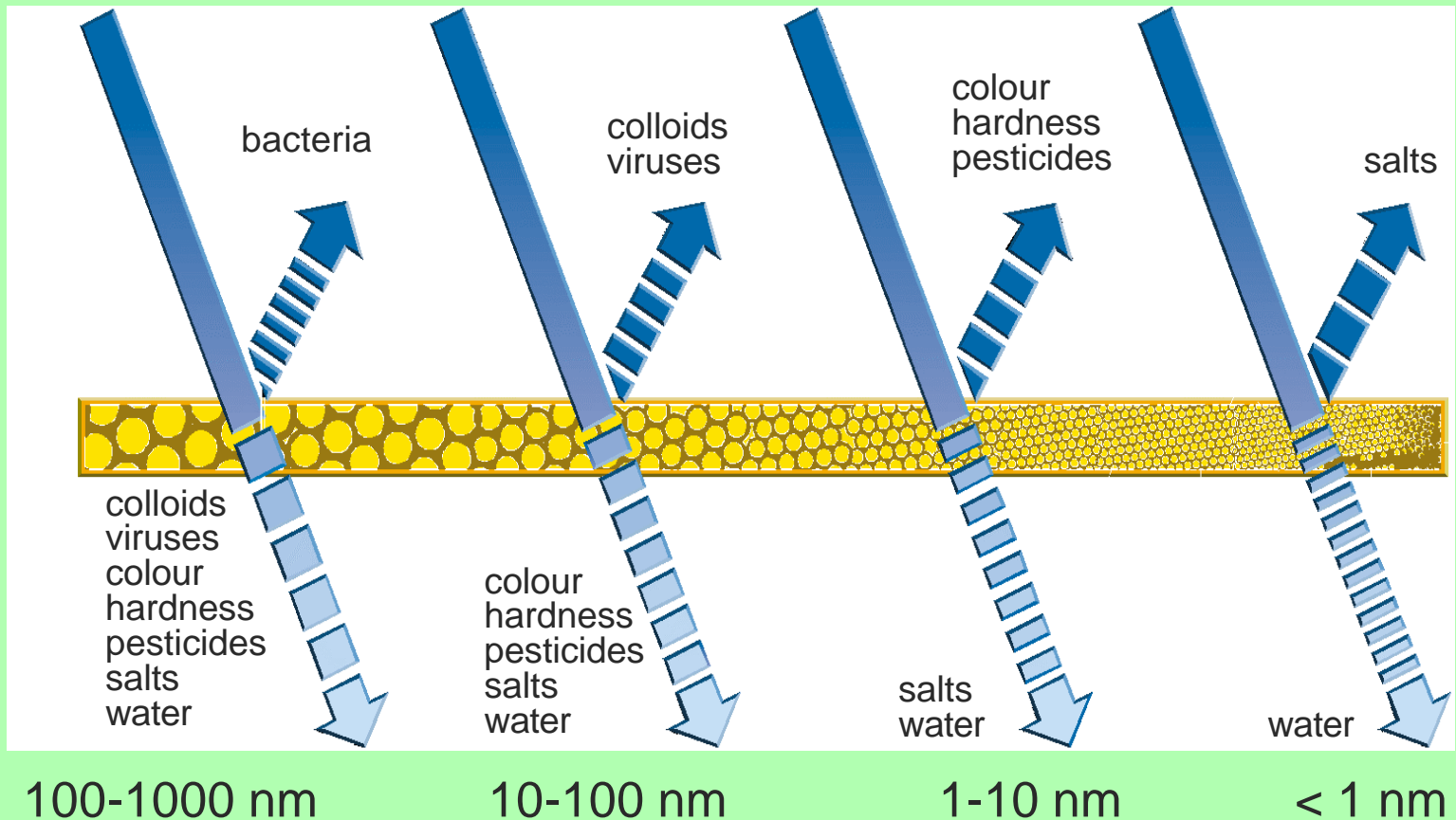
FILTRATION TECHNIQUES

Micro Filtration

Ultra Filtration

Nano Filtration

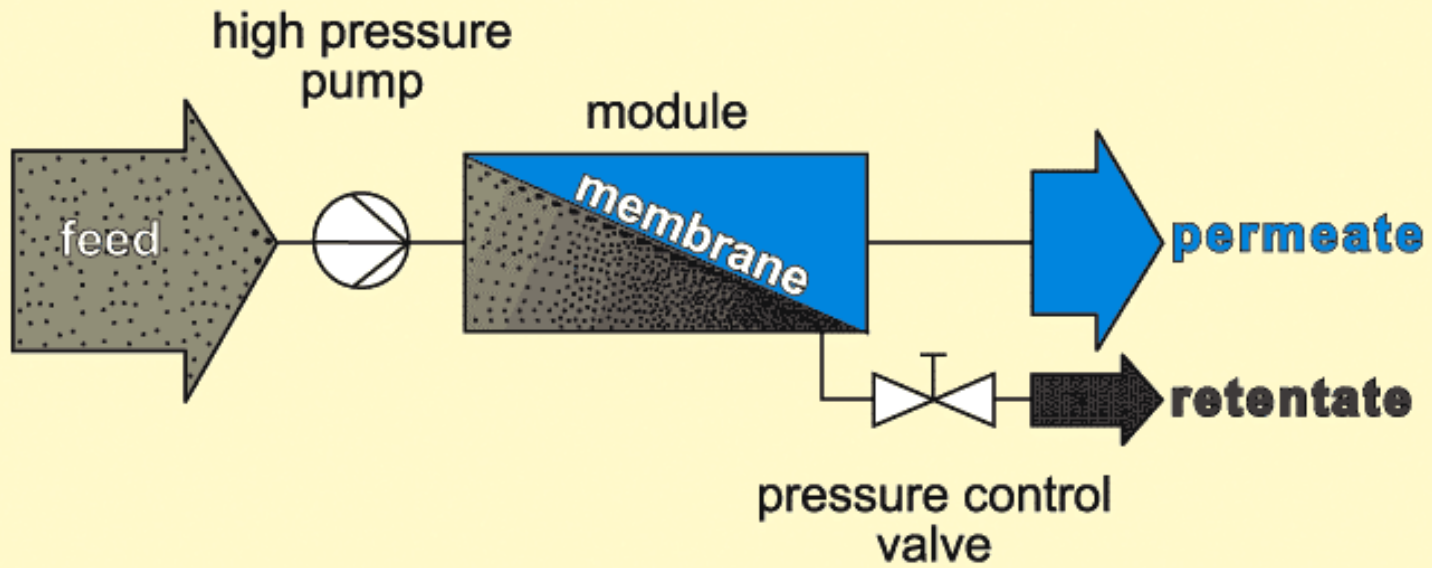
Reverse Osmosis



REVERSE OSMOSIS

The residual contaminants of biologically treated leachate might be concentrated by reverse osmosis and evaporation.

The overall treatment process and specially the residue discharge are expensive.



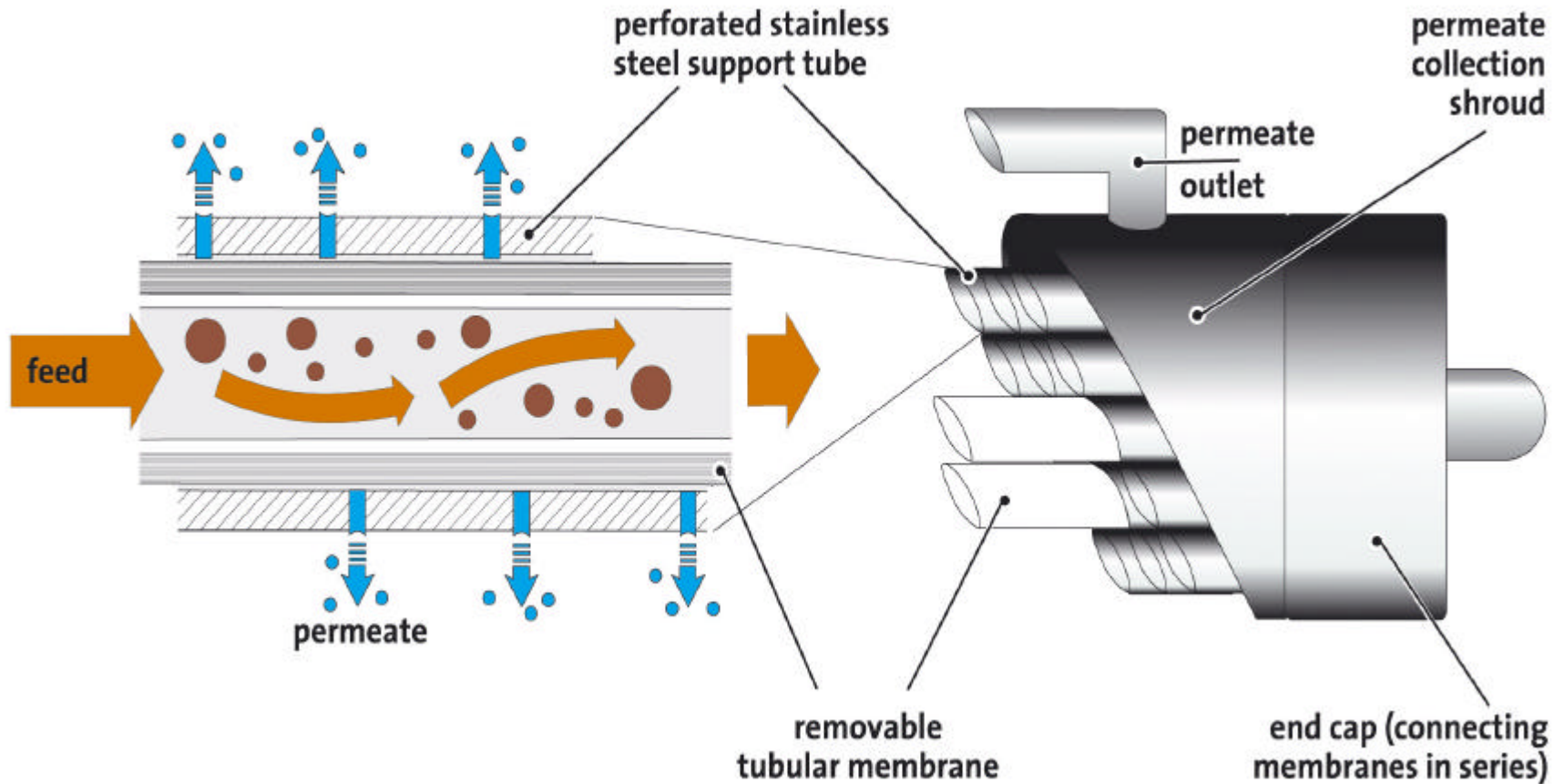
TYPICAL RO PLANT PERFORMANCE IN LEACHATE PURIFICATION

| PARAMETER | leachate | permeate I | permeate II | rejection in% |
|---------------------------------|-----------------|-------------------|--------------------|----------------------|
| pH-value | 7.7 | 6.8 | 6.6 | |
| el. conduct.* | 17,250 | 382 | 20 | 99.9 |
| COD in mgO₂/l | 1,797 | < 15 | < 15 | > 99.2 |
| ammonium mg/l | 366 | 9.8 | 0.66 | 99.9 |
| chloride mg/l | 2,830 | 48.4 | 1.9 | 99.9 |
| sodium mg/l | 4,180 | 55.9 | 2.5 | 99.9 |
| heavy metals # | 0.25 | < 0.005 | < 0.005 | > 98 |

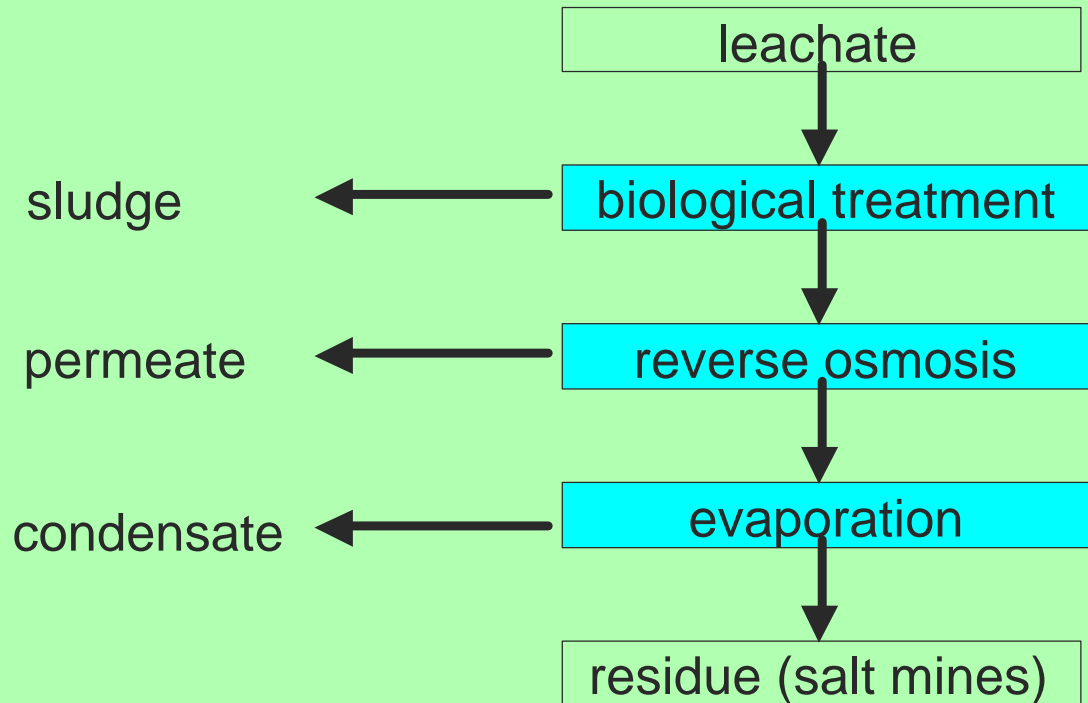
* = $\mu\text{S/cm}$ # = mg/l



TUBULAR MODULE



Scheme of the reverse osmosis - evaporation plant



REVERSE-OSMOSIS



EVAPORATION PLANT



COLORS



influent NF

effluent NF



| <i>parameter</i> | <i>unit</i> | <i>cleaning efficiency [%]</i> | | |
|---------------------|-------------|--------------------------------|------------------|-----------------------|
| | | <i>reverse osmosis</i> | | <i>nanofiltration</i> |
| | | <i>one-stage</i> | <i>two-stage</i> | |
| <i>ammonium</i> | [mg / l] | 86,0 – 94,0 | 98,0 – 99,6 | ca. 5 – 40 |
| <i>nitrate</i> | [mg / l] | 86,0 – 90,0 | 98,0 – 99,0 | ca. 5 – 40 |
| <i>conductivity</i> | [µS / cm] | 84,0 – 95,0 | 97,4 – 99,7 | ca. 5 – 40 |
| <i>COD</i> | [mg / l] | 90,0 – 96,0 | 99,0 – 99,8 | ca. 85 – 95 |
| <i>BOD</i> | [mg / l] | 90,0 – 96,0 | 99,0 – 99,8 | ca. 80 – 90 |
| <i>AOX</i> | [mg / l] | 90,0 – 94,0 | 99,0 – 99,6 | ca. 80 – 90 |
| <i>heavy metals</i> | [mg / l] | 86,0 – 90,0 | 98,0 – 99,0 | ca. 80 – 95 |

Dahm, Kollbach, Gebel: „Sickerwasserreinigung“, 1994



Adsorption

Treatment of old leachate or of biological pretreated leachate

Low efficiency with volatile acids due to polarity

Removal of low biodegradable compounds (Residual-COD and AOX)

Removal of heavy metals

High efficiency at low pH

Removal of fulvic acids (MW: 100-10000)

Difficult removal of humic acids due to too high MW

- powder or granular Activated Carbon
- Rotating Reactors
- Columns
- 500mg COD/g AC
- Regeneration of Loaded Activated Carbon is necessary



FLOCCULATION AND PRECIPITATION

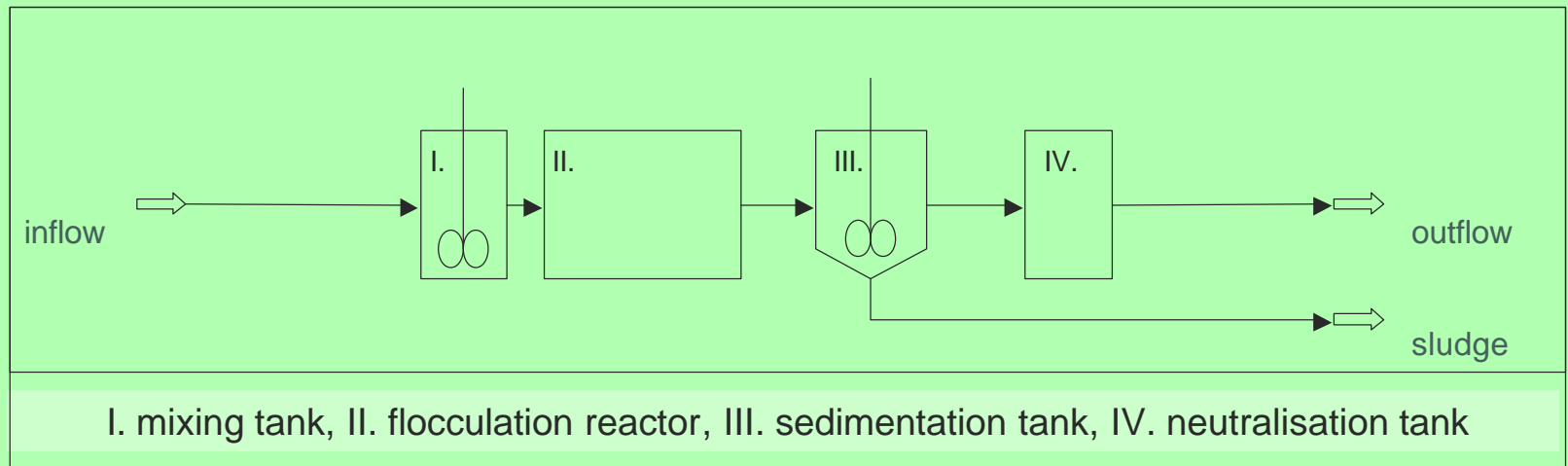
Precipitation: Formation of insoluble compounds by means of addition of auxiliary substances: the equilibrium is shifted towards the insoluble form

Removal of Metals

Flocculation: suspended, colloidal or emulsified substances become destabilized and precipitate

Removal of Organic Compounds

Up to 50% COD Reduction



nach Dahm, Kollbach, Gebel: „Sickerwasserreinigung“, 1994

ESTIMATIONS FOR COSTS OF COMBINATIONS FOR LEACHATE TREATMENT IN RELATION TO THE CAPACITY

| Treatment (combination) | costs for | costs for |
|--------------------------|---|--|
| | small capacity (< 10m ³ /h) [€/m ³] | high capacity (> 10m ³ /h) [€/m ³] |
| Biology + Membrane Sep. | 9 - 30 | 7 - 15 |
| Biology + Chemical Oxid. | 18 - 50 | 9 - 30 |
| Biology + Active Carbon | 3 - 25 | 1 - 10 |
| Biology + Flocc./Precip. | 3 - 30 | 3 - 15 |
| Biology + Reverse Osmos. | 5 - 25 | 3 - 8 |
| Evaporation | | 6 - 24 |



POSSIBLE TREATMENT COMBINATIONS

